A review of hydrochory (seed dispersal by water) with implications for riparian rehabilitation

Jade Hyslop and Sam Trowsdale

School of Environment, University of Auckland, Private Bag 92019, Auckland, New Zealand. (Corresponding author: s.trowsdale@auckland.ac.nz)

Abstract

Working alongside natural regeneration processes is an important part of an ecosystem approach to long-term, self-sustaining, riparian rehabilitation. Hydrochory (seed dispersal by water) is a key influence on such rehabilitation and has received limited attention. This review of hydrochory suggests it is the relationships between hydrology, fluvial geomorphology and seed biology which determine seed deposition and germination. In-stream geomorphic and hydraulic diversity are important for both seed transport and seed retention in fluvial systems. Woody debris and flotsam are important for longdistance seed dispersal, especially for seeds without specialised floating phenology. Lateral connectivity between the channel and floodplains, as well as timing of seed release, for example with receding flood flows, influence whether seeds are transported to sites with favourable germination conditions. Not all hydrochory is desirable, such as the spread of invasive plant species in New Zealand. It is clear that much more attention should be given to hydrochory to better align riparian management with natural ecosystem processes and improve the longterm outcomes of riparian rehabilitation.

Keywords

hydrochory; seed dispersal; fluvial processes; ecosystem approach; riparian rehabilitation

The role of hydrochory in an ecosystem-based approach to riparian rehabilitation

Substantial time, energy and money have been invested in riparian re-vegetation and weed eradication in the past thirty years (Palmer et al., 2005). Given these investments and the importance of riparian vegetation both as an amenity and for river health (Smaill et al., 2011), there is inherent pressure for rehabilitation projects to be successful (Palmer et al., 2005; Palmer and Allan, 2006). Palmer et al. (2005) argued that most riparian schemes are nowhere near as successful in the long-term as they should be, due to plant dieback and weed invasions; and there are a number of challenges in ensuring effective implementation and successful results (Brierley et al., 2006; Fryirs and Brierley, 2009). Groves et al. (2007) demonstrated that a key reason for reduced planting success is a limited understanding of long-term riparian sustainability, particularly for areas that are located in isolated pockets and are susceptible to weed infestation. It is clear that rivers and streams continue to degrade, despite an increase in rehabilitation efforts (Palmer and Allan, 2006).

The advent of ecological perspectives prompted a paradigm shift in river conservation through the 1970s and 1980s. During this period papers stressed the adversity of human impacts on river systems,

and promoted research and investment in riparian and river rehabilitation (e.g., Behnke, 1978; Kauffman and Krueger, 1984; Glade, 2003; Brierley et al., 2005). A move to more ecosystem-based management is evident in approaches in which systems are viewed as continuously changing entities, influenced by a complex array of controls, including social processes (Glade, 2003; Brierley et al., 2005). Nevertheless, in practice, many riparian revegetation projects follow traditional horticultural approaches to riparian planting (e.g., personal communication with Chris Ferkins Project Twin Streams Waitakere City Council, 2010). The approach stems from a 'command and control' ideology, which emphasises human control over nature and involves a high level of human investment in rehabilitation. Each seedling is selected and placed in a given site and a typical measure of success is survival rate. This approach has many disadvantages, including high monetary costs and human effort and may not lead to a naturally functioning ecosystem.

Differing approaches to riparian vegetation have caused debate over how to manage riparian planting. With growing awareness of the critical relationship between vegetation dynamics and river forms and processes emerging from the 1980s onward, researchers have questioned how interdisciplinary research, in terms of a whole-of-system approach, might be used to enhance the results of river and riparian rehabilitation (e.g. Allen *et al.*, 2011).

A number of researchers have suggested that natural seed dispersal and deposition would be useful to incorporate into riparian rehabilitation (e.g. Groves *et al.*, 2007; Gurnell *et al.*, 2008; Goodson et al., 2003). Seeds can be dispersed across the landscape in a variety of ways, including birds and other animals, gravity, wind and water. Hydrochory – seed dispersal by water – is an important mechanism, particularly in

riparian environments and may influence seed bank structuring and riparian composition (Nilsson et al., 1991; Johansson et al., 1996; Andersson et al., 2000; Goodson et al. 2002). Fenner and Thompson (2005) state that all riparian species are likely to be dispersed by water to some extent. Seeds are released into water where they can be dispersed downstream at great distances from parent trees or onto higher elevations during flooding, depending on flow magnitude (Naiman et al., 1988). Skoglund (1990) found an average of 35,000 seeds m⁻² in a 10-cm layer of drift material that had been deposited onto floodplains along the lower Dalalven River, Sweden. Hydrochory may also remobilise species already dispersed by other mechanisms (Bennett and Simon, 2004).

Although water is an obvious dispersal vector in riparian environments, it has received limited attention in literature compared to the other dispersal mechanisms. Riparian rehabilitation that is mindful of the role of hydrochory in seed dispersal will ultimately achieve better long-term rehabilitation results. Clearly the key processes that drive water dispersal must be understood. As a step towards incorporating hydrochory in successful riparian rehabilitation, this paper reviews the current knowledge of seed dispersal by water and its associated processes, focusing on links between hydrology, seed biology and fluvial geomorphology. Discussion is provided on ways that knowledge of these processes and their interactions might be applied.

A review of research on hydrochory

Botanists have recognised hydrochory as an important mechanism of seed dispersal since Darwin (Goodson *et al.*, 2002), although relatively few studies dealing specifically with water dispersal were documented until the latter half of the 20th century. Willson

et al. (1990) examined the spectrum of seed dispersal mechanisms, but specific seed adaptations to water dispersal were deemed too rare to warrant more than a mention. Research has grown rapidly in the past decade or so. Nilsson et al. (2010) reported an almost exponential increase in research reports from 1984-1999 to 2000-2010, and more recently hydrochory has been shown to be a very effective mechanism for moving seeds.

Hydrochory research has implications for flow regulation, restoration, climate change and the spread of native and invasive species. Studies have shown that water dispersal is important for transporting and depositing freshly produced seeds (Cellot et al., 1998; Thebauld and Debussche, 1991; Merritt and Wohl, 2002; Boedeltje et al., 2003; Goodson et al., 2003; Tabacchi et al., 2005), remobilising seeds (Goodson et al. 2003; Pettit & Froend 2001), structuring riparian plant communities (Andersson et al., 2000; Johansson et al., 1996; Nilsson et al., 1991; Goodson et al., 2002; Tabacchi et al., 2005; Groves et al., 2009) and maintaining high riparian diversity (Nilsson et al., 1994; Johansson et al., 1996; Andersson et al., 2000). Of course, hydrochory can disperse native species through a river course and spread invasive weed species, e.g., Salix spp. commonly known as Willow, and *Usnea* spp. or Old Man's Beard, which are both weeds in New Zealand (Groves et al., 2009). Hydrochory can thus be both beneficial and detrimental to efforts to manage riparian plants.

The influence of water dispersal on riparian vegetation composition is debated in literature. Some studies have measured seed quantity and vegetation composition and found organic flotsam along rivers to be richer in species than the wind-dispersed seed rain, which suggests that flotsam in streams and rivers is an important mechanism for dispersing seeds (Merrit and Wohl, 2002).

Hughes and Cass (1997) ascertained that water dispersal has an important influence on vegetation structure by comparing propagule species with standing vegetation species. In their study of a floodplain forest, newly deposited seeds transported in flotsam during flooding were found to contribute new species to the deposition site (40 seeds were deposited by flood waters, of which only 14 were present in the standing vegetation). This difference in species type supports the idea that river transport plays a role in structuring patterns of vegetation.

Research by Andersson et al. (2000) supported the conclusion that hydrochory is important for riparian species richness. Wooden cubes were used as a proxy for seeds and released in the free-flowing Vindel River, Sweden. The number of cubes deposited was found to correlate with high above-ground species richness. Additionally, it was found that seed floating ability was not important for vegetation composition because flotsam could transport seeds during regular spring floods. Most recently, Nilsson et al. (2010) suggested that although water has the potential to colonise sites that other dispersal vectors may not reach, this depends on the interplay between other factors, such as the timing of dispersal and the environmental conditions of a deposition site necessary for successful establishment.

Parolin (2005) categorised hydrochory into three forms: (1) dispersal by water currents on the surface, (2) dispersal by water currents on the bottom of a channel, and (3) dispersal by rain falling on a plant. These dispersal types (particularly the first two) play a role in long-distance dispersal, which is important for landscape connectivity and for maintaining and extending species populations at the landscape scale (Merrit and Wohl, 2006). Nilsson *et al.* (2010) suggested that water can disperse seeds to greater distances than other dispersal mechanisms.

Clearly, increased knowledge of long-distance dispersion through river corridors can improve understanding of how landscapes are colonised by plants.

A number of processes influence hydrochorous seed dispersal and deposition patterns, but many of the process relationships and seed responses remain unclear. Merrit and Wohl (2002) identified three processes: '(1) the hydrological regime during seed release and transport, which is determined by the timing and magnitude of peak flow and the rate and direction of change in discharge, (2) channel morphology and (3) the phenology of seed release as it relates to the hydrologic regime'. Researchers tend to agree that once seeds enter the fluvial system, the local hydrological regime is the most important in terms of determining dispersal distance and deposition location (e.g., Schneider and Sharitz, 1986; Edwards et al., 1994; Walker et al., 1995; Danvind and Nilsson, 1997; Poff et al., 1997; Middleton, 2000; Andersson et al., 2000; Pettit et al., 2001; Merritt and Wohl, 2002; Nilsson et al., 2002; Boedeltje et al., 2003; Hampe, 2004; Edwards et al., 1994; Riis and Sand-Jensen, 2006; Richter et al., 1997). However, the way that the hydrology interacts with fluvial geomorphology, stream hydraulics, and seed biology together determine the final location of water-dispersed seeds. These controls are dynamically adjusted, meaning that a change in one will often produce a change in the others (Brierley and Fryirs, 2000; Gurnell et al., 2002; Montgomery et al., 2003). Francis (2006) has suggested that more advanced approaches in riparian management are limited by the absence of research that considers interrelationships between hydrology, geomorphology and ecology.

The influence of flow regime on seed dispersal

Walker *et al.* (1995) described the flow regime as 'the maestro that orchestrates pattern and process'. Ecologically significant aspects of flow regime include flow magnitude, flow variability, rate of flow change, magnitude and frequency of flow conditions, and flow predictability (Richter *et al.*, 1996).

Flow velocity, in particular surface flow velocity, plays a key role in determining seed entrainment and dispersal distance. Merrit and Wohl (2002) found significant differences in seed deposition between units with low surface flow (e.g., pools, slack waters) and units with high surface flow (e.g., straight reaches, cut banks, riffle-run sequences, cascades, rapids, waterfalls). Groves et al. (2009) found variation of flow velocity between reaches to be the major driver underlying differences in seed suspension and deposition. Chambert and James (2009) found that different flow velocities along the dispersal pathway led to preferential sorting of seed species type. Johansson and Nilsson (1993) and Andersson et al. (2000) suggested that small increases in flow velocity result in an almost exponential increase in the dispersal distance of seed mimics. Fischer et al. (1979) described the relationships between particle dispersal and velocity. An increase in velocity results in an increase in shear stress. This in turn reduces surface flow dead zones along the bank, which limits the ability of the bank to trap seeds, and leads to extended seed dispersal. As a result of differences in shear stress, seeds move faster through the central streamline than along channel banks (Fischer et al., 1979).

Studies on sediment dispersal processes have proved useful to understand seed dispersal. The relationship between sediment and seed deposition was first observed by Nilsson *et al.* (1993) in their study on northern European rivers. Goodson *et al.* (2003) found

a positive relationship between hydrochorous seed deposition and sediment deposition at different elevations along the Dover River, England, and Cerdà and Garcia-Fayos (2002) reported in laboratory experiments that seeds (<50mg) were suspended at the same flow threshold as similar-sized sediment particles.

Groves et al. (2007) described similarities in seed and sediment movement. Seeds that float are generally transported with washload sediments (e.g., fine clays), which move with similar velocities and directions to surface flow, although seeds can settle out of the washload as they become waterlogged. Surface flow velocity varies laterally and longitudinally through the river channel in relation to channel roughness, which is controlled by riparian vegetation cover, large woody debris and channel sinuosity (Groves et al., 2007). The location and concentration of washload sediments and floating seeds also vary throughout the river channel. Washload sediments and floating seeds move most quickly through the channel thalweg and settle out in areas with still surface flow, such as pools and slack water and on bar surfaces, alongside channel banks and floodplains. As a result of this geomorphic and hydraulic diversity, floating particles are challenging to model at small reach and geomorphic unit scales (Meade et al., 1990).

Floating seeds can become waterlogged and sink over time. Dispersal pathways for sinking seeds are analogous to those for suspended or bedload material (Groves *et al.*, 2007). These materials are entrained into suspension when the stream power threshold is met, usually at higher than normal flow stage, and can be deposited onto areas that are infrequently inundated and disturbed providing opportunity to germinate.

A large number of sediment dispersal models have been developed; they vary in complexity, including their input variables/processes, the scale to which they are applied, and their inherent assumptions.

A comprehensive review of sediment transport models, including detailed descriptions of several common models, is provided by Merrit *et al.* (2003), who concluded that although there are clear advantages in the use of sediment dispersal models, more work is needed to generate simple, physically realistic models that will be of most practical use to river managers. The accuracy of spatial modelling required in sediment dispersal models remains an issue, as do a lack of datasets for calibration (Merrit *et al.*, 2003), which are similar issues facing seed dispersal models.

As flow regime is very important for seed dispersal patterns, a number of studies have focused on seed dispersal in regulated rivers (e.g., Andersson *et al.*, 2000; Jansson *et al.*, 2000; Jansson *et al.*, 2006). The majority of studies in regulated rivers have noted the positive contributions that hydrochory can make to riparian biodiversity but, as we know, hydrochory can also disperse invasive weeds.

Regulated flow regimes have been found to constrain seed dispersal. Jansson et al. (2000) found riparian composition to vary between adjacent impoundments in a regulated river, as a result of a forced uneven distribution of floating propagules. Merrit and Wohl (2006) found higher species richness in reaches downstream of reservoirs compared to upstream, suggesting that seeds from an upstream catchment area are combined with seeds from local sources along freeflowing river reaches, whereas seeds along regulated reaches are predominantly from local sources. Andersson et al. (2000) found dam impoundments to have much weaker currents than free-flowing rivers, limiting dispersal distances.

As dispersal distances can be limited by regulated flows, the probability of seeds reaching suitable sites for germination is also reduced (Francis, 2006). Studies have demonstrated the role of flooding in

connecting otherwise disjunct populations (e.g., Vogt *et al.* 2004; Schneider and Sharitz, 1986). Overall, seed deposition patterns during floods remain poorly understood. Of course, seeds need to be available for floods to disperse the seeds.

Implications of the connections between flow regime and hydrochory for riparian rehabilitation

Flow velocity is particularly important for determining seed entrainment, transport and deposition. It also leads to preferential sorting of seeds along the river margin. Seeds are likely to be deposited in low-flow surface units, such as pools and slack water, while units with higher surface flow, such as straight reaches, cut banks and riffle-run sequences, can increase dispersal distance. It is important to maintain a diversity of low and high surface flows, which is a function of hydrology and channel roughness, so that seeds can be dispersed to downstream rehabilitation sites but can also be retained in the system for remobilisation in flood flows. Environmental flows in regulated rivers may increase seed dispersal distances and the likelihood of seeds reaching sites suitable for germination.

Biological influences on hydrochory – traits of seed phenology and life history

The timing of seed releases determines when seeds will enter the fluvial system. Relationships between the timing of seed releases and flow regimes are clearly important for establishing dispersal distances, pathways, and suitable sites for germination. Kubitzki and Ziburski (1994) demonstrated that the synchronisation of fruiting with predictable floods is the most important adaptation to hydrochory. For water to be an effective means of transport, seed fall needs to coincide with favourable flow conditions, such as receding

floodwaters when soil is exposed (White, 1979).

Merrit and Wohl (2002) found a significant relationship between flow regime and timing of seed release, which resulted in distinctive deposition patterns along channel margins. Pettit and Froend (2001) studied four dominant species found along the Blackwood and Ord Rivers, Australia, and found seed fall to coincide with the end of the rainy season. Seeds were thus released during high flows and high flow recessions and were therefore deposited onto sediments ideal for germination (bare and moist). This adaptation by seed phenology is found with Populus and Salix species in the central USA (Bradley and Smith, 1986) and in the southeastern USA (Schneider and Sharitz, 1988), and was reported by Estabrook et al. (1982) to be fairly common among some species.

The anatomical characteristics of seeds, such as their size and shape, might make them prone to particular dispersal vectors. A study by Andersson *et al.* (2000) found that *Helianthus annuus* (sunflower) seeds were deposited in very similar ways to wooden cubes (22mm in size) and suggested that depositional patterns across a range of seed sizes may be consistent. Particular seed morphologies or life history traits aid dispersal by water (Howe and Smallwood, 1982; Johansson and Nilsson, 1993; Malanson, 1993). A general question which remains largely unanswered is whether different seed floating abilities influence riparian distribution patterns.

A number of studies have looked at the effects of seed buoyancy traits on transport pathways, with varying results. van den Broek *et al.* (2005) linked increasing seed buoyancy of vegetation zones with increasing inundation rates along wetland ecosystems in The Netherlands and Germany. The study illustrated that seed buoyancy is important for determining the distribution of a range of species in wetland systems, which are inundated for a long period of time each year.

In contrast, a study by Danvind and Nilsson (1997) hypothesised that seed floating ability would correlate with long-distance dispersal, yet found that species distribution was not significantly correlated with variation in seed buoyancy and that high floating ability is not necessary for effective dispersal. One obvious explanation is that woody debris and flotsam can trap seeds and increase seed dispersal distances from the parent plant, regardless of floating ability (Johansson and Nilsson, 1993). This was illustrated in a dispersal study by Nilsson et al. (2002) in which wooden cubes, used as seed mimics, were mostly stranded on riverborne litter deposits. Nilsson et al. (1994) suggested that spring floods can transport seeds a long way in a short period of time (the spring flood peak moves along a 230 km stretch in 2.5-0.64 days in the freeflowing Vindel River, Sweden). This allows seeds even with brief floating abilities to be dispersed over long distances. Therefore, seed buoyancy may not be as important in rivers with high flotsam loadings and in freeflowing rivers with unregulated high (spring) flood peaks.

Nilsson et al. (2002) found that seed buoyancy may influence vegetation composition at the reach scale; linking to hydraulic variability between reaches. Species with long-floating propagules were found more along river lakeshores and tranquil reaches and short-floating species found more along turbulent reaches. Overall, the importance of seed buoyancy in structuring vegetation communities seems to depend on river type, reach-scale variability, and associated flow paths.

Fenner and Thompson (2005) argued that it is difficult to be certain of any particular dispersal mechanism based on phenological attributes alone. Although many species possess buoyancy characteristics, seeds without these traits can still float or become suspended in the water. Danvind and Nilsson (1997) suggested that the influence

of buoyancy is complicated by multiple mechanisms of dispersal.

Dispersal distance is an important factor in determining which native (or invasive) stands upstream contribute to downstream plant rehabilitation. Seeds have higher potential of retaining viability if they are deposited in drift materials along channel margins; otherwise they may become waterlogged, sink to the channel bed and lose viability (Hughes and Cass, 1997; Andersson *et al.*, 2000). Andersson *et al.* (2000) found that river banks which received the most drift material, including seeds, were the most species rich.

Implications of the connections between plant biology and hydrochory for riparian rehabilitation

The connections between timing of seed fall, flood flows and geomorphic availability are important for hydrochorous seed transport, deposition and germination. The best time for seeds to enter a stream or river is during receding floodwaters so that seeds can be transported long distances and deposited onto elevated floodplain surfaces. Floodplains are often left coated with moist and fertile sediments following flood flows, which are particularly favourable conditions for germination. They are also infrequently inundated or disturbed, providing an opportunity for plants to establish. Flood flows are clearly important for riparian rehabilitation to be self-sustaining, as is high lateral connectivity between a river and its floodplains. Furthermore, the fact that organic flotsam and woody debris aid seed transportation means that seeds do not require specialised floating ability to be dispersed by water.

Fluvial geomorphology and hydrochory

The *quantity* of seeds and the *quality* of the dispersal of each seed combine to determine

'disperser effectiveness', which was defined by Schupp (1993) as 'the contribution a dispersal makes to the future reproduction of a plant'. This is largely dependent on the biotic and abiotic conditions that characterise the sites on which seeds become deposited (Schupp, 1993; Dennis *et al.*, 2007). Fluvial geomorphology (including substrate characteristics, channel shape and size, geomorphic unit assemblages, floodplain attributes) provides the physical template upon which seeds may become deposited, germinated, buried, reworked, or remobilised.

Species tolerance to particular hydrogeomorphic conditions, along with competition by other species, are key factors driving successful species germination (Bendix and Hupp, 2000). Geomorphology of the deposition surface plays a key role in plant growth and riparian species composition.

Different geomorphic unit types have characteristic micro-environmental conditions. There are dynamic relationships between hydrology and fluvial geomorphology, involving substrate destruction and creation, which work together to influence patterns and directions of seed dispersal, and create groupings of geomorphic units (Bendix and Hupp, 2000). Analysis of units can be used to interpret the fluvial processes that formed them, which is a function of the available energy and sediment characteristics (Brierley and Fryirs, 2005).

Individual geomorphic units have inherent hydrological characteristics, for example, bars and benches are inundated under bankfull flow stage, whilst floodplains are inundated in overbank flow stage. Geomorphic units and groupings are easily identified in the field and can be used in management to select appropriate plant species for the site. Seeds may also be stored on river beds, thus channel bed geometry is important for preferential seed storage (Gurnell *et al.*, 2007).

A benchmark study by Merrit and Wohl (2002) used an experimental flume study to

explore whether seed dispersal and deposition along river margins could be predicted on the basis of hydrological regime, geomorphic features, and timing of release. More seeds were found deposited in areas of reduced velocities, which related to geomorphic and hydraulic units including eddies, flow expansions, point bars, pool margins, and slack water. Conversely, fewer seeds were found deposited on units relating to high flow velocities such as cut banks, flow constrictions, islands and straight margins. This demonstrates the importance of relationships between flow regime and geomorphic unit type and the potential to use geomorphic units as a proxy for predicting species deposition. The study suggests that low surface-flow velocities are important for retaining seeds in the river system, where they can be remobilised in flood flows and deposited onto sites more suited to plant germination.

Geomorphic groupings vary longitudinally, laterally, and vertically as a result of catchment position and related controls. Vegetation patches form in relation to different geomorphic groupings at varying scales, which include: '(1) the landscape scale (contrasts in physical environmental characteristics between reaches), (2) the reach scale (contrasts between within-reach landforms such as bars and floodplain surfaces) and (3) the bar scale (contrasts between patches of varying sedimentary or topographical characteristics within the same bar)' (Francis, 2006).

Depending on the particular scale of enquiry, different geomorphic controls, processes and features will relevant. Vogt et al. (2004) recommended that high-resolution sampling measures (reach or geomorphic unit which link to hydraulics) are required for meaningful results when studying hydrochorous seed transport. Geomorphic assemblages clearly differ according to river type.

Most riparian species require bare soils to become established, as they tend to be pioneer species that do not compete well with later succession species (Francis, 2006). However, site suitability for germination is variable, as geomorphic forms and their spatial arrangements continuously adjust (Schupp, 1993). Even so, the presence or absence of particular geomorphic characteristics within a reach, at a given time, is clearly significant for determining patterns of seed dispersal. Connectivity is also important, particularly between rivers and their floodplains, so that seeds can be transported to suitable sites for germination (Bornette et al., 1998; Tockner and Ward, 1999).

Vegetation, sediment and litter transported by rivers can collect on preferential geomorphic units (Goodson et al., 2001). Changxing et al. (1999) found organic matter containing numerous seeds accumulated on concave benches and bars along the regulated lower Dee River, UK. Following the scouring of these surfaces in post-flood events, the newly bare sites provided ideal surfaces for seed germination. Deposited seeds may also form part of the seed bank (Goodson et al., 2001).

Andersson *et al.* (2000) correlated the stranding patterns of an artificial diaspore (wooden cubes) with environmental variables in the free-flowing Vindel River, northern Sweden, and found some conflicting results. Rapids was the only predictive variable correlating to number of stranded cubes at a site. Correlations were not significant between riverbank height, width, or vegetation cover. This could be explained by the spring flows, which vary in elevation and duration between years. Additionally, weather conditions influence different wind patterns and currents, which affect dispersal patterns.

Despite the importance of these connections, the role of fluvial geomorphology in seed dispersal has been largely overlooked in hydrochory literature. Much more could be made of the potential for fluvial geomorphic insight to assist in riparian rehabilitation, particularly at the unit scale.

The geomorphic substrate provides the surface upon which seeds can be deposited, remobilised and/or germinate. Geomorphic units and groupings of units involve substrate destruction and creation due to hydrology and substrate type. Diversity of geomorphic units is important to provide a range of surfaces for plants with different hydrogeomorphic tolerances to germinate. It is likely that seeds will deposit on low-flow units, which is important for seed retention in the river system. These surfaces are frequently reworked and seeds may become remobilised in flood flows and transported to sites more suitable for germination.

Geomorphic units and their groupings can be used to estimate inundation frequency where flood data does not exist. This could be useful in selecting plants most suited to the given site. For example, the New Zealand natives *Cortaderia richardii* (toe-toe), *Phormium tenax* (harakeke, New Zealand flax) and *Carex secta* (purei) tolerate water inundation and could grow on geomorphic surfaces that flood periodically (Ledgard and Henley, 2009).

Incorporating hydrochory into ecosystem-based riparian rehabilitation

Current understandings of the key processes involved in structuring riparian communities, including seed supply, dispersal, deposition and germination are not being fully utilised in riparian rehabilitation (e.g., Holmes *et al.*, 2005). Hydrochory is an important seed dispersal mechanism in riparian environments (Groves *et al.*, 2007; Gurnell *et al.*, 2008; Nilsson and Svedmark, 2002) however there is very little guidance in scientific literature

on how knowledge of hydrochory might be incorporated into riparian rehabilitation plans. This paper suggests that an understanding of the key processes that drive hydrochory, including hydrology, fluvial geomorphology and seed biology, can be used in riparian rehabilitation to aid natural regeneration processes.

Petts and Amaros (1996) listed three principles for any ecosystems approach to river management. Firstly is the need for a catchment-scale approach. This is a reoccurring message in rehabilitation literature (e.g., Brierley and Fryirs, 2000; Allan et al., 2011). Rehabilitation strategies should reflect upstream catchment conditions, e.g., consideration of native vegetation sources in the upstream catchment, catchment modifications and how these have affected river flows. Longitudinal connectivity is clearly important. Management at an individual site or reach might provide shortterm benefits, but catchment scale planning will be required for long-term success (Petts and Amaros, 1996).

Secondly, lateral exchanges at the landwater interface are important for sustaining long-term functionality of river systems. Thirdly, there is a need to view the natural range of variability of rivers systems, involving hydrological, geomorphological and ecological interactions. It makes sense that riparian and river rehabilitation plans need to be adapted depending on river type and catchment modifications.

Implicit in this message is the need to 'know your catchment', so that river and riparian rehabilitation strategies can be adapted to work with the natural conditions and characteristics of a particular site. This might involve several strands of work: historical review of the river catchment to determine if and where the river has been modified; ascertaining channel-floodplain connectivity (frequency of bankfull flow stage/overbank

flow stage); mapping geomorphic features; and mapping riparian plants to locate native /invasive plants in the catchment (i.e., what seeds are likely to be dispersed to the rehabilitation site).

A conceptual model (Fig.1) illustrates how interactions between flow stage, geomorphic diversity (meaning the range of geomorphic units such as bars, benches, floodplains) and seed phenology influence hydrochorous seed dispersal and deposition in riparian ecosystems. The value of understanding these process interactions can be demonstrated by considering their influence on seed deposition and implications for rehabilitation in two conceptual scenarios – intact and modified catchments.

In intact catchments with lots of native riparian vegetation and high in-stream geomorphic diversity, seeds that enter the stream during low flow stage may be retained in units with slow surface flow such as pools and slack water, and on geomorphic units such as bars and benches. Seeds retained in these slow-flow units may be remobilised in high flood events and transported onto floodplains, which can provide favourable conditions for seed germination. An input of woody debris and plant flotsam materials into the stream from intact riparian zones in the catchment will aid seed transport and increase seed dispersal distances. A rehabilitation strategy might consider protecting native seed sources in the catchment particularly those up-stream. Minimal active planting should be required. The New Zealand focus might involve weeding and site monitoring.

In a modified catchment with low instream geomorphic and hydraulic diversity and little input of woody debris and plant flotsam, seeds that enter the stream or river during low flow stage have less chance of being retained and/or deposited and are more likely to be transported through the fluvial system. If flood flows are restricted (e.g., stop

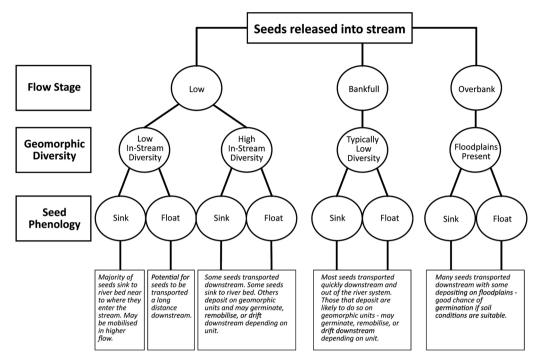


Figure 1 – Interactions between flow stage, geomorphic diversity and seed phenology and their influence on hydrochorous seed dispersal and deposition.

banks, river channeling, river bed scouring), this reduces the lateral connectivity between the channel and floodplains, with less opportunity for hydrochorous seeds to reach floodplains and germinate. A key strategy in working towards a more functioning and natural system may be to reconnect riparian zones that have been isolated from the river (system?) (Nilsson and Svedmark, 2002). Native seed sources could be protected and native riparian vegetation in the upstream catchment could be promoted. Propagule dispersal (local soil seed banks or intact patches of vegetation in the catchment) have been found to be critical for natural recovery (Holmes et al., 2005).

In highly modified catchments, native propagule dispersal is likely to be low. In these cases planting may be required both in the upstream catchment and at key rehabilitation sites. Critically, the planting should take into account fluvial geomorphology, substrate conditions and likely inundation frequency. Some rehabilitation sites may lack environments that are suitable for seed germination and might require new strategies designed to work within a new environmental setting.

In highly modified rivers, floodplainchannel lateral connectivity and high flow events may need to be re-established. Instream geomorphic diversity could also be reestablished to generate hydraulic variability in the channel and create sites for seed deposition. Nilsson and Svedmark (2002) described passive management strategies as those which 'let the river do the work', such as reconstructing natural geometry by reconnecting floodplains to river flows, for example by lowering embankments and weirs. In the lower Missouri River vegetation responded quickly to the reconnection with its floodplain, with the recovery of perennial and invasive plants following flood scouring (Nilsson and Svedmark, 2002).

With any rehabilitation efforts, the complex relationship between processes means that there will always be uncertainty about where seeds will be deposited. However, knowledge of how key processes interact can be used to plan rehabilitation strategies and link the ideas of seeds entrainment and deposition with fluvial geomorphology and hydrology for more desirable outcomes.

Conclusions

A review of hydrochory suggests the subject has been largely ignored until recently when research has grown rapidly. Hydrochory has been shown to be important for transporting, depositing and remobilising seeds, structuring plant communities and maintaining riparian diversity. Improved understanding and utilisation of hydrochory and the processes that drive it (importantly hydrology, seed biology and fluvial geomorphology) may lead to more self-sustaining riparian rehabilitation practices.

There are complex interactions between hydrology, seed biology and fluvial geomorphology. Surface flow is an important transport vector as many seeds float and those without floating abilities can be transported in rivers or streams in woody debris or flotsam. Seeds will collect in areas of slow flow, such as pools and slack water and will be transported downstream in fast flowing zones, such as riffle-run sequences. Hydraulic diversity is clearly important for both seed transport and seed retention in a river system.

The timing of seed release interacts with flow regime to influence dispersal pathways and distance. Floodwaters are especially important, as seeds can be transported and deposited onto floodplain surfaces, which may be left bare and moist following inundation. This provides favourable germination conditions. In modified river catchments, flood flows have been re-instated to enable long-distance seed dispersal and improve channel-floodplain connectivity.

Fluvial geomorphology provides the physical template, upon which seeds can be deposited, buried, reworked, remobilised, or germinate. There are clear hydrology and fluvial geomorphology process relationships that involve substrate creation and destruction. Geomorphic diversity (e.g., bars, benches, banks, floodplains) is important for seed deposition and germination, as different plants have different hydro-geomorphic tolerances. Vegetation patches are known to form in relation to these different geomorphic groupings and at varying spatial scales. Riparian rehabilitation must take into account the interaction between native and invasive plant species. Geomorphic landforms, particularly substrate conditions and inundation frequency, should be considered in the selection of suitable protection and planting sites (if planting is necessary). A hydrologic and geomorphic appraisal should be part of any riparian management efforts. Integrated approach to riparian rehabilitation that consider processbased interactions across river contexts and at different scales is needed to improve the implementation of hydrochory and the longterm success of riparian rehabilitation.

References

Allen, W.; Fenemor, A.; Kilvington, M.; Harmsworth, G.; Young, R.; Deans, N.; Horn, C.; Phillips, C.; Montes de Oca, O.; Ataria, J.; Smith, R. 2011: Building collaboration and learning in integrated catchment management: the importance of social process and multiple engagement approaches. *Journal of Marine and Freshwater Research* 45(3): 525-539.

Andersson, E., Nilsson, C.; Johansson, M.W. 2000: Plant Dispersal in Boreal Rivers and its Relation to the Diversity of Riparian Flora. *Journal of Biography* 27: 1095-1106.

- Behnke, R.J. 1978: Grazing and the Riparian Zone: impact on aquatic values. Lowland Rover and Stream Habitat in Colorado, a Symposium, Oct. 4-5,1978, Greeley, Colorado. University of Northern Colorado,. Audubon Council, Soil Conservation Society, and Bureau of Land Management: 126-132.
- Bendix, J., Hupp, C.R. 2000: Hydrological and geomorphological impacts on riparian plant communities. *Hydrological Processes* 14: 2977-2990.
- Bennett, S.J.; A. Simon, (eds.) 2004: Riparian Vegetation and Fluvial Geomorphology. Water and Science Application, vol. 8, American Geophysical Union, Washington, DC, 282 pp.
- Boedeltje, G.; Bakker, J.P.; Bekker, R.M., van Groenendael, J.M.; Soesbergen, M. 2003: Plant dispersal in a lowland stream in relation to occurrence and three specific life-history traits of the species in the species pool. *Journal* of Ecology 91: 855-866.
- Bornette, G.; Amoros, C.; Lamourous, N. 1998: Aquatic plant diversity in riverine wetlands: the role of connectivity. *Freshwater Biology 39(2)*: 267-283.
- Bradley, C.E.; Smith, D.G. 1986: Plains cottonwood recruitment and survival on a prairie meandering river floodplain, Milk River, Southern Alberta and Northern Montana. *Canadian Journal of Botany 64*: 1433-1442.
- Brierley, G.J.; Fryirs, K.A. 2000: River Styles, A Geomorphic Approach to Catchment Characterization: Implications for River Rehabilitation in Bega Catchment, New South Wales, Australia. *Environmental Management* 25(6): 661-679.
- Brierley, G.J.; Fryirs, K.A. 2005: Geomorphology and River Management: Application of the River Styles Framework. Blackwell Publishing, Massachusetts, USA, 416 pp.
- Brierley, G.J.; Hillman, M.; Fryirs, K. 2006: Knowing your place: an Australasian perspective on catchment-framed approaches to river management. *Australian Geographer 37* (2): 131-145.
- Cellot, B.; Mouillot, F.; Henry, C.P. 1998: Flood drift and propagule bank of aquatic macrophytes in a riverine wetland. *Journal of Vegetation Science 9*: 631-640.

- Cerdà, A.; Garcia-Fayos, P. 2002: The influence of seed size and shape on their removal by water erosion. *Catena* 48: 293-301.
- Chambert, S.; James, C.S. 2009: Sorting of seeds by hydrochory. *River Research and Applications* 25: 48-61.
- Changxing, S.; Petts, G.; Gurnell, A.1999: Bench development along the regulated, lower River Dee, UK. *Earth Surface Processes and Landforms* 24(2): 135-149.
- Danvind, M.; Nilsson, C. 1997: Seed floating ability and distribution of alpine plants along a northern Swedish river. *Journal of Vegetation Science* 8(2): 271-276.
- Dennis, A.J.; Schupp, E.W.; Green, R.A.; Westcott, D.A. 2007: Seed Dispersal: Theory and its Application in a Changing World. Cooperative Research Centre for Tropical Rainforest Ecology and Management (CSIRO), Biddles Ltd, UK.
- Edwards, A.; Wyatt, R. Sharitz, R, 1994: Seed buoyancy and viability of the wetland milkweed *Asclepias perrenis* and an upland milkweed, *Asclepias exaltata. Bulletin of the Torrey Botanical Club 121(2)*: 160-169.
- Estabrook, G.F.; Winsor, J.A.; Stephenson, A.G.; Howe, H.F. 1982: When Are Two Phenological Patterns Different? *Botanical Gazette 143(3)*: 374-378.
- Fenner, M.; Thompson, K. 2005: *The Ecology of Seeds*. Cambridge University Press, Cambridge, UK.
- Fischer, H.B.; List, E.J.; Koh, R.C.Y.; Imberger, J.; Brooks, N.H. 1979: *Mixing in Inland and Coastal Waters*. Academic Press, New York.
- Francis, R.A. 2006: Allogenic and autogenic influences upon riparian vegetation dynamics. *Area 38*: 453-464.
- Fryirs, K.; Brierley, G. 2009: Naturalness and Place in River Rehabilitation. *Ecology and Society 14(1)*: 20.
- Glade, T. 2003: Landslide occurrence as a response to land use change: a review of evidence from New Zealand. *Catena* 51(3-4): 297-314.
- Goodson, J.M., Gurnell, A.M.; Angold, P.G.; Morrissey, I.P. 2001: Riparian seed banks: structure, process and implications for riparian management. *Progress in Physical Geography 25*: 301-325.

- Goodson, J.M.; Gurnell, A.M.; Angold, P.G.; Morrissey, I.P. 2002: Riparian seed banks along the lower River Dover, UK: their structure and ecological implications. *Geomorphology* 47(1): 45-60.
- Goodson, J.M., Gurnell, A.M.; Angold, P.G.; Morriseey, I.P. 2003: Evidence for hydrochory and the deposition of viable seeds within winter flow-deposited sediment: the River Dove, Derbyshire, UK. River Research & Environmental Law 18: 113-121.
- Groves, J.; Caitcheon, G.; Norris, R. Williams, D. 2007: Prediction of fluvial seed dispersal and long-term sustainability of riparian vegetation using sediment transport processes. In: Proceedings of the 5th Australian Stream Management Conference, Australian Rivers: Making a Difference, Wilson, A.L., Dehaan, R.L.; Watts, R.J.; Page, K.J.; Bowmer, K.H.; Curtis, A.(eds.). Charles Sturt University, Thurgoona, New South Wales.
- Groves, J.H.; Williams, D.G.; Caley, P.; Norris, R.H.; Caitcheon, G. 2009: Modelling of floating seed dispersal in a fluvial environment. *River Research and Applications* 25(5): 582-592.
- Gurnell, A.M. 2007: Analogies between mineral sediment and vegetative particle dynamics in fluvial systems. *Geomorphology* 89(1-2): 9-22.
- Gurnell, A.M.; Petts, G.E. 2002: Island-dominated landscapes of large floodplain rivers, a European perspective. *Freshwater Biology* 47(4): 581-600.
- Gurnell, A.; Goodson, J.; Thompson, K.; Clifford, N.; Armitage, P. 2007: The riverbed: a dynamic store for plant propagules? Earth Surface Processes and Landforms 32: 1257-1272.
- Gurnell, A.; Thompson, T.; Goodson, J.; Moggridge, H. 2008: Propagule deposition along river margins: linking hydrology and ecology, *Journal of Ecology 96*: 553-565.
- Hampe, A. 2004: Extensive hydrochory uncouples spatiotemporal patterns of seedfall and seedling recruitment in a 'bird-dispersed' riparian tree. *Journal of Ecology 92*: 797-807.
- Holmes, P.M.; Richardson, D.M.; Esler, K.J.; Witkowski, E.T.F.; Fourie, S. 2005: A decisionmaking framework for restoring riparian zones degraded by invasive alien plants in South Africa. South African Journal of Science 101: 553-564.

- Howe, H.F.; Smallwood, J. 1982: Ecology of Seed Dispersal. *Annual Review of Ecology and Systematics 13*: 201-228.
- Hughes, J.W.; Cass, W.B. 1997: Pattern and Process of a floodplain forest, Vermont, USA: predicted response of vegetation to perturbation. *Journal of Applied Ecology 34(3)*: 594-612.
- Jansson, R.; Nilsson, C.; Renofalt, B. 2000: Fragmentation of Riparian Floras in Rivers with Multiple Dams. *Ecology* 81(4): 899-903.
- Jansson, R.; Žinko, U.; Merrit, D.; Nilsson, C. 2005: Hydrochory increases riparian plant species richness: a comparison between a freeflowing and a regulated river. *Journal of Ecology* 93: 1094-1103.
- Johansson, M.E.; Nilsson, C. 1993: Hydrochory, population dynamics and distribution of the clonal aquatic plant Ranunculus lingua. *Journal of Ecology 81(1)*: 81-91.
- Johansson, M.E., Nilsson, C.; Nilsson, E. 1996: Do rivers function as corridors for plant dispersal? *Journal of Vegetation Science* 7(4): 593-598.
- Kauffman, J.B.; Krueger, W.C. 1984: Livestock Impacts on Riparian Ecosystems and Streamside Management Implications: A Review. *Journal of Range Management* 37(5): 430-438.
- Kubitzki, K.; Ziburski, A. 1994: Seed Dispersal in Flood Plain Forests in Amazonia. *Biotropica* 26(1): 30-43.
- Ledgard, N.; Henley, D. 2009: Native plant establishment along riparian margins of the Sherry River, Motueka catchment: 'Best bet' guidelines, Technical report produced by Scion, Christchurch for Landcare Research, NZ.
- Malanson, G.P. 1993: *Riparian Landscapes*. Cambridge University Press, Cambridge, UK.
- Meade, R.H.; Yuzyk, T.R.; Day, T.J. 1990: Movement and storage of sediment in rivers of the United States and Canada. In: Surface Water Hydrology, Geological Society of America: 255-280.
- Merrit, D.M.; Wohl, E. 2006. Plant dispersal along rivers fragmented by dams. *River Research and Applications 22*: 1-26.
- Merrit, D.M.; Wohl, E. 2002: Processes governing hydrochory along rivers: hydraulics, hydrology, and dispersal phenology. *Ecological Applications* 12(4): 1071-1087.

- Merrit, W.S.; Letcher, R.A.; Jakeman, A.J. 2003: A review of erosion and sediment transport models. *Environmental Modelling & Software* 18: 761-799.
- Middleton, B. 2000: Hydrochory, seed banks, and regeneration dynamics along the landscape boundaries of a forested wetland. *Plant Ecology* 146: 169-184.
- Montgomery, D.R.; Collins, B.D.; Abbe, T.B.; Buffington, J.M. 2003: Geomorphic Effects of Wood in Rivers. In: *The Ecology and Management of Wood in World Rivers, American Fisheries Society Symposium, vol. 37.* Gregory, S.; Boyer, K.L. (eds.), American Fisheries Society: 1-10.
- Naiman, R.J.; Decamps, H.; Pastor, J.; Johnston, C.A. 1988: The Potential Importance of Boundaries to Fluvial Ecosystems. *Journal of* the North American Benthological Society 7: 289-306.
- Nilsson, C.; Grelsson, G. 1990: The effects of litter displacement on riverbank vegetation. *Canadian Journal of Botany 68*: 735-741.
- Nilsson, C.; Svedmark, M. 2002: Basic Principles and Ecological Consequences of Changing Water Regimes: Riparian Plant Communities. Environmental Management 30(4): 468-480.
- Nilsson, C.; Andersson, E.; Merrit, D.M.; Johansson, M.E. 2002: Differences in Riparian Flora between Riverbanks and River Lakeshores Explained by Dispersal Traits. *Ecology* 83(10): 2878-2887.
- Nilsson, C.; Brownn, R.L.; Jansson, R.; Merrit, D.M. 2010: The role of hydrochory in structuring riparian and wetland vegetation. *Biological Reviews 85(4)*: 1-24.
- Nilsson, C.; Ekblad, A, Dynesius, M.; Backe, S.; Gardfjell, M.; Carlber, B.; Hellqvist, S.; Jansson, R. 1994: A comparison of species richness and traits of riparian plants between a main river channel and its tributaries. *Journal of Ecology 82*: 281-295.
- Nilsson, C.; Gradfiell, M.; Grelsson, G. 1991: Importance of hydrochory in structuring plant communities along rivers. *Canadian Journal of Botany* 69: 2631-2633.
- Nilsson, C.; Nilsson, E.; Johansson, M.E.; Dynesius, M.; Grelsson, G.; Xiong, S.; Jansson, R.; Danvind, M. 1993: Processes structuring riparian vegetation. In: *Current Topics in Botanical Research*, Menon, J. (ed.), Council of Scientific Research Integration, Triviandrum, India: 419-431.

- Palmer, M.A.; Bernhardt, E.S.; Allan, D.; Lake, P.S.; Alexander, G.; Brooks, S.; Carr, J.; Clayton, S.; Dahm, C.N.; Follstad Shah, J.; Galat, D.L.; Loss, S.G.; Goodwin, P.; Hart, D.D.; Hassett, B.; Jenkinson, R.; Kondolf, G.M.; Lave, R.; Meyer, J.L.; O'Donnell, T.K.; Pagano, L.; Sudduth, E. 2005: Standards for Ecologically Successful River Restoration. *Journal of Applied Ecology* 42: 208-217.
- Palmer, M.A.; Allan, J.D. 2006: Restoring Rivers: The work has begun but we have yet to determine what works best. *Issues in Science and Technology 22(2)*: 40-48.
- Parolin, P. 2005: Ombrohydrochory: Rainoperated seed dispersal in plants - With special regard to jet-action dispersal in Aizoaceae. *Flora 201*: 511-518.
- Pettit, N.E.; Froend, R.H. 2001: Availability of seed recruitment of riparian vegetation: a comparison of a tropical and a temperate ecosystem in Australia. *Australian Journal of Botany* 49: 515-528.
- Petts, G.; Amaros, C. 1996: *The Fluvial Hydrosystem*. Chapman and Hall Ltd, London.
- Poff, L.; Allan, J.D.;, Bain, M.B.; Karr, J.R.; Prestegaard, K.L.; Richter, B.; Sparks, R.; Stromberg, J. 1997: The Natural Flow Regime: A paradigm for river conservation and restoration. *Bioscience* 47: 769-784.
- Richter, B.D.; Baumgartner, J.V.; Powell, J.; Braun, D.P. 1996: A Method for Assessing Hydrologic Alteration within Ecosystems. *Conservation Biology 10*: 1163-1174.
- Richter, B.D.; Baumgartner, J.V.; Wigington, R.; Braun, D.P. 1997: How much water does a river need? *Freshwater Biology* 37: 231-249.
- Riis, T.; Sand-Jenson, K. 2006: Dispersal of plant fragments in small streams. *Freshwater Biology* 51: 274-286.
- Schneider, R.L.; Sharitz, R.R. 1986: Hydrochory and Regeneration in a Bald Cypress Water Tupelo Swamp Forest. *Ecology* 69: 1055-1063.
- Schupp, E.W. 1993: Quantity, quality and the effectiveness of seed dispersal by animals. *Vegetatio 107/108*: 15-29.
- Skoglund, S.J. 1990: Seed dispersing agents in two regularly flooded river sites. *Canadian Journal of Botany 68*: 754-760.

- Smaill, S.J.; Ledgard, N.; Langer, E.R.; Henley, D. 2011: Establishing native plants in a weedy riparian environment. *New Zealand Journal of Marine and Freshwater Research* 45(3): 357-367.
- Tabacchi, E.; Planty-Tabacchi, A.M.; Roques, L.; Nadal, E. 2005: Seed inputs in riparian zones: implications for plant invasion. *River Research and Applications 21*: 299-313.
- Thebauld, C.; Debussche, M. 1991: Rapid invasion of *Fraxinus ornus* along the Herault River system in southern France: the importance of seed dispersal by water. *Journal of Biogegraphy 18*: 7-12.
- Tockner, K.; Ward, J.V. 1999: Biodiversity along riparian corridors. *Archiv für Hydrobiologie 3*, *Supplement, 115*: 293-310.
- van den Broek, T.; van Diggelen, R.; Bobbink, R. 2005: Variation in seed buoyancy of species in wetland ecosystems with different flooding dynamics. *Journal of Vegetation Science 16*: 579-586.

- Vogt, K.; Rasron, L.; Jensen, K. 2006: Seed deposition in drift lines during an extreme flooding event: Evidence for hydrochorous dispersal? Basic and Applied Ecology 7: 422-432.
- Vogt, K.; Rasran, L.; Jensen, K. 2004: Waterborne seed transport and seed deposition in a small river-valley in Northern Germany. *Flora* 199: 377-388.
- Walker, K.F.; Sheldon, F.; Puckridge, J.T. 1995: A perspective on dryland river ecosystems. Regulated Rivers: Research & Management 11: 85-104.
- White, P.S. 1979: Pattern, Process and Natural Disturbance in Vegetation. *Botanical Review* 45: 229-299.
- Willson, M.F.; Rice, B.L.; Westoby, M. 1990: Seed dispersal spectra: a comparison of temperate plant communities. *Journal of Vegetation Science* 1:547-562.